Surface water quality and cropping systems sustainability under a changing climate in the Upper Mississippi River Basin


Abstract: Agricultural nonpoint source pollution is the main source of nitrogen (N) and phosphorus (P) in the intensely row-cropped Upper Mississippi River Basin (UMRB) stream system and is considered the primary cause of the northern Gulf of Mexico hypoxic zone according to the US Environmental Protection Agency. A point of crucial importance in this region is how intensive corn (Zea mays L.)-based cropping systems for food and fuel production can be sustainable and coexist with a healthy water environment, not only under existing climate conditions but also under a changed climate in the future. To address this issue, a UMRB integrated modeling system has been built with a greatly refined 12-digit subbasin structure based on the Soil and Water Assessment Tool (SWAT) water quality model, which is capable of estimating landscape and in-stream water and pollutant yields in response to a wide array of alternative cropping and/or management strategies and climatic conditions. The effects of the following four agricultural management scenarios on crop production and pollutant loads exported from the cropland of the UMRB to streams and rivers were evaluated: (1) expansion of continuous corn across the entire basin, (2) adoption of no-till on all corn and soybean (Glycine max L.) fields in the region, (3) substitution of the traditional continuous corn and corn–soybean rotations with an extended five-year rotation consisting of corn, soybean, and three years of alfalfa (Medicago sativa L.), and (4) implementation of a winter cover crop within the baseline rotations. The effects of each management scenario were evaluated both for current climate and a projected midcentury (2046 to 2065) climate from a General Circulation Model (GCM). All four scenarios behaved similarly under the historical and future climate, generally resulting in reduced erosion and nutrient loadings to surface water bodies compared to the baseline agricultural management. Continuous corn was the only scenario which resulted in increased N pollution while no-till was the most environmentally effective and able to sustain production at almost the same levels. Rye (Secale cereale L.) cover crop within the fallow period was also effective in reducing erosion and both sediment-bound and soluble forms of nutrients. The results indicated that alternative management practices could reduce sediment, N, and P exports from UMRB cropland by up to 50% without significantly affecting yields. Results for the climate change scenario showed that the effectiveness of the management scenarios was strongly linked to the reduced water availability predicted under the future climate, which assisted in mitigating pollutant transport, although with a small loss of production.

Keywords: agricultural management scenarios—climate change—corn based systems—nonpoint pollution—Soil and Water Assessment Tool—Upper Mississippi River Basin

Overenrichment of nutrients constitutes a major problem in many streams and rivers in the United States. In addition to local effects, transport of these nutrients contributes to environmental problems such as eutrophication in downstream lakes, bays, and estuaries, and is primarily responsible for hypoxia in the Gulf of Mexico (USEPA 2000). The Mississippi River/Gulf of Mexico Watershed Nutrient Task Force (2008) established a goal to reduce the size of the hypoxic zone in the Gulf of Mexico to 5,000 km² (1,900 mi²). This will require substantial reductions in nutrient loadings from the Mississippi/Atchafalaya River Basin and especially from its most upstream and intensively cultivated part, the Upper Mississippi River Basin (UMRB), which forms the western part of the Corn Belt region of the United States. Within this large area, tradeoffs between the interdependent goals of sustainable biofuel production, food production, and water resources can have significant implications for commodity groups, individual producers and other stakeholders in the region.

Within this context, physically based hydrological models can be used to evaluate socioeconomic and environmental impacts of agricultural management scenarios. In order to reliably address what-if scenarios for future agriculture, however, the impacts of future climate change should also be accounted for. The Soil and Water Assessment Tool (SWAT) water quality model (Arnold et al. 1998; Williams et al. 2008) has proven to be an effective tool for evaluating agricultural management simulations for complex landscapes and varying climate regimes, including future climate projections (Gassman et al. 2007, 2013; Douglas-Mankin et al. 2010; Tuppard et al. 2011).
The Soil and Water Assessment Tool has been applied in several previous UMRB based assessments including studies focused on calibration and validation approaches (Srinivasan et al. 2010; Santhi et al. 2012), climate change effects on hydrology and water quality (Jha et al. 2006, 2013; Wu et al. 2012), and evaluation of land use or best management practice (BMP) scenarios (Rabotyagov et al. 2010; Secchi et al. 2011; USDA NRCS 2012; Demissie et al. 2012). However, none of these studies investigated the impact of climate on the efficiency or environmental consequences of alternative management scenarios. Here, we investigate the impacts of both present climate and a projected future climate on the following land use and BMP scenarios for the UMRB: (1) conversion of all cropland to a continuous corn (Zea mays L.; C-C) rotation, (2) adoption of no-till (NT) on all cropland areas, (3) extended rotation of corn–soybean (Glycine max L.; C-S) and C-C with alfalfa (Medicago sativa L.), and (4) the adoption of a winter cover crop (rye [Secale cereal L.]) within rotations of corn and soybean. In addition, a greatly refined SWAT subbasin delineation approach is introduced here; this approach allows for improved linkages to climate data, due to the SWAT structure that requires climate data to be input to a given subbasin from the closest climate station. Refined subbasin structure allows input of downscaled, bias-corrected General Circulation Model (GCM) projections across a dense grid overlaid on the UMRB study region.

Materials and Methods

Watershed Description. The UMRB is a headwater basin of the Mississippi River and extends from Lake Itasca in Minnesota to just north of Cairo, Illinois, above the confluence with the Ohio River (Srinivasan et al. 2010). It covers approximately 492,000 km² (190,000 mi²), including large parts of Illinois, Iowa, Minnesota, Missouri, and Wisconsin (figure 1). The area is referred to as Region 07 by the US Geological Survey (USGS) at a 2-digit watershed scale and is further comprised of 131 USGS 8-digit basins (USGS 2012). The average annual UMRB rainfall within the last 4 decades was 900 mm (35.4 in), ranging from 600 mm (23.6 in) to 1,200 mm (47.2 in) across the basin with values generally decreasing from east to west. Cropland consists mainly of corn–soybean rotations and occupies 50% of the total UMRB area, with 75% of the land under gentle slopes. According to USEPA SAB (2007), 43% of the nitrate (NO₃) load and 26% of the total phosphorus (P) load delivered to the Gulf of Mexico came from the UMRB during 2001 to 2005, even though the UMRB covers only 15% of the total Mississippi River drainage area. The mean annual flow of the Mississippi River at Grafton, Illinois (figure 1), is 3,500 m³ s⁻¹ (123,600 ft³ s⁻¹), where the river loads for nitrogen (N) and P have been measured as 500,000 t (551,000 tn) and 30,000 t (33,060 tn) respectively (USGS 2013). More than two-thirds of the total N load occurs as NO₃-

Soil and Water Assessment Tool Model Description. SWAT is a river basin model with extensive worldwide use (Gassman et al. 2007). It was developed by the USDA in collaboration with Texas A&M University (Williams et al. 2008) and is continuously upgraded with improved versions and interfaces. A recent release of SWAT version 2012 (SWAT2012) in combination with the ArcGIS (version 10.1) SWAT (ArcSWAT) interface (SWAT 2013) were used in this study. In SWAT, a watershed is typically delineated into subbasins and subsequently into Hydrologic Response Units (HRUs), which represent homogeneous combinations of land use, soil types, and slope classes in each subbasin (but are not spatially identified within a given subbasin). However, a dominant HRU approach can also be used in which no further delineation of subbasins occurs; i.e., a given subbasin is synonymous with a single HRU (which was the method used in this study). The physical processes associated with water and sediment movement, crop growth, and nutrient cycling are

Figure 1
The 12-digit Upper Mississippi River Basin (UMRB) and the calibration points along Mississippi River and its tributaries.
modelled at the HRU scale; runoff and pollu-
tants exported from the different HRUs are ag-
gregated at the subbasin level and routed
downstream. Simulation of the hydrology is
separated into the land and the routing
phase of the hydrological cycle, while sedi-
ment yields are estimated with the Modified
Universal Soil Loss Equation (MUSLE;
Neitsch et al. 2009). SWAT simulates both
N and P cycling, which are influenced by
specified management practices. Both N and
P are divided in the soil into two parts, each
associated with organic and inorganic N and
P transport and transformations. Agricultural
management practices can be simulated with
specific dates and by explicitly defining the
appropriate management parameters for
each HRU. In-field conservation practices
such as contour farming, strip-cropping, ter-
races and residue management are simulated
with changes to model parameters that repre-
sent cultivation patterns (Arabi et al. 2008).

The Soil and Water Assessment Tool Upper
Mississippi River Basin Parameterization.

Key data layers that were incorporated for
building the UMRB SWAT model included
climate, soil, land use, and topographic and
management data sources. Topography was
represented by a 30 m (98.43 ft) digital ele-
vation model (USGS 2013) and was used in
ArcSWAT to calculate landscape parameters
such as slope and slope length. As previously
noted, a greatly refined delineation scheme
has been incorporated into the current model,
which consists of using subbasin boundaries
that are coincident with the USGS 12-digit
watersheds instead of the coarser 8-digit
watersheds, which have been used in pre-
vious SWAT UMRB studies (figure 2).
The area of a 12-digit watershed is typically
4,000 to 16,000 ha (10,000 to 40,000 ac),
compared with around 200,000 to 400,000
ha (500,000 to 1,000,000 ac) for an 8-digit
watershed. Historic daily precipitation, and
maximum and minimum temperatures were
obtained from the National Climatic Data
Center (NCDC-NOAA 2012) and were
input to the model from a total of more
than 1,000 climate stations across the study
region. Wind speed, relative humidity, and
solar radiation data, required for the esti-
mation of potential evapotranspiration (ET)
using the Penman-Monteith method, were
generated internally in SWAT using the
model's weather generator.
The landuse layer of the UMRB model
was created by using the USDA NASS
Cropland Data Layer (CDL) datasets (USDA
NASS 2012) in combination with the 2001
National Land Cover Data (NLCD 2001).
This approach included the overlay of three
years of CDL datasets in order to create
crop rotations used in the region, similar
to the approach reported by Srinivasan et
al. (2010) for the UMRB. Two-year rota-
tions of C-S rotation dominated within
the overall agricultural land portion of the
region with a smaller fraction managed with
C-C rotation. Soil characteristics were rep-
resented by the USDA 1:250,000 State Soil
Geographic (STATSGO) soil data (USDA
NRCS 2013). The resolution of these data
was rather coarse with approximately 600
soil types lying within the UMRB. Thus,
we overlaid land use and soils on each of the
5,729 subbasins in ArcSWAT and selected
the dominant land use type and soil occupy-
ing each subbasin. Therefore, the number
of HRUs in this study was equal to the num-er of subbasins; i.e., one 12-digit subbasin
is one HRU. This approach resulted in a
slight (<2%) increase of the total cropland
area compared to the original land use map,
with forest being also slightly increased,
while other land cover types were reduced
accordingly to maintain the sum of all land
types equal to the total area of the basin.
Minor rotations such as corn–corn–soybean
or corn–soybean–wheat (Triticum aestivum)
were eliminated in this process, which com-
prised less than 5% of the cropland area in
any of the 12-digit subbasins.
Estimates of possible locations where
subsurface tiles are used to drain soils, a key
conduct of NO₃ to surface waters, were based
on areal county-level estimates compiled by
Sugg (2007). Estimates at the county level
were first aggregated at the 8-digit level with
the use of geographic information system
(GIS) applications in order to have the same
spatial reference with available fertilizer
and tillage data. Tile drains were first assigned
to the agricultural subbasins (12-digit water-
sheild) within each 8-digit watershed with
slopes lower than 2% and with poorly
drained soils (hydrologic groups D or C),
and subsequently to low-slope, hydrologic
group B soil if needed. All tile drains were
simulated with the following assumptions:
a depth of 1,200 mm (3.94 ft), the time
to drain a soil to field capacity (24 hours),
and the amount of time required to release
water from a drain tile to a stream reach (72
hours), which are the SWAT DDRAIN,
TDRAIN, and GDRAIN input parameters,
respectively (see Neitsch et al. [2009] for fur-
ther description of these input parameters).
Spatial representation of various tillage
types (conventional, reduced, mulch, and
no-till) were incorporated in the modeling
system using data reported by Baker (2011),
who compiled estimates of the distributions
of different tillage types at the 8-digit water-
shed level using survey data collected by
the Conservation Technology Information

Regional estimates of the distribution of other conservation practices were not publicly available at the time of this study. To address this deficiency, we used a proxy approach that was based on information provided in the Conservation Effects Assessment Project (Duriancik et al. 2008) UMRB study (USDAs NRCS 2012). They reported that more than 45% of the cropland in the UMRB had at least one in-field conservation practice (terrace, strip-cropping, or contouring), while highly erodible land was managed to a much greater extent compared to less erodible areas. In our model, the conservation practices were likely to be present in all the HRUs due to their relatively large areas (12-digit subbasins). Therefore, we simulated the effect of in-field conservation practices on erosion control in all of them by reducing the management (P) factor of the MUSLE (Neitsch et al. 2009), which was the major parameter that governed the representation of all such practices in the model (Arabi et al. 2008). Similarly, we reduced the slope length, but did not modify HRUs with slopes less than 2.3% as in those HRUs estimated erosion was inversely correlated with slope length (Arabi et al. 2008). We specified higher reductions of the management P factor in high-sloping agricultural HRUs and slight reductions in low sloping ones. These adjustments of the P factor had also the purpose of calibrating the model for sediment based on model comparisons with observations available from various USGS monitoring stations (USGS 2013). Adjustment of curve numbers (CNs), which are additionally used to represent such practices (Arabi et al. 2008), was not implemented because the CNs served as one of the key parameters for calibrating the hydrological UMRB model (see next subsection), and the reduced CN values that resulted from the flow calibration during the final 15-year period coincided with expanded adoption of conservation tillage and other conservation practices in the UMRB region.

Fertilizer (including manure) application rates were calculated based on recent estimates at the 8-digit level obtained from the Nutrient Use Geographic Information System (NuGIS) for the United States (IPNI 2010). However, problems were encountered in applying these data in the current modeling system due to uncertainty in the fertilizer sales data used in NuGIS and other factors. Thus, statewide averages computed from the NuGIS data were used in the present study, resulting in annual average N and P rates applied to cropland that ranged between 117 to 156 kg ha\(^{-1}\) yr\(^{-1}\) (103 to 137 lb ac\(^{-1}\) yr\(^{-1}\)) and 25 to 34 kg ha\(^{-1}\) yr\(^{-1}\) (22 to 30 lb ac\(^{-1}\) yr\(^{-1}\)), respectively, with N applied only to corn. For hay and pastureland we used the autofertilization routine of SWAT by setting a 70 kg N ha\(^{-1}\) yr\(^{-1}\) (62 lb ac\(^{-1}\) yr\(^{-1}\)) as maximum limit.

Monthly streamflow data obtained from 12 UMRB USGS stations (figure 1) were used for calibrating the model (USGS 2013), with the most downstream station located at Grafton, Illinois. These data were obtained for 1975 to 2010, with the most recent 14-year period used for calibration and the rest for validation. In-stream sediment, NO\(_3\) -N, organic N, and organic and mineral P data were available for most of these stations on a monthly basis for similar or shorter time-periods. Calibration of river sediment and nutrient yields was also conducted for all the locations with available data after incorporating N and P loads from thousands of point sources across the region (Maupin and Ivahnenko 2011; Dale Robertson, personal communication).

Model Performance and Evaluation.

The hydrologic calibration of the UMRB model was conducted with the use of the SWAT Calibration and Uncertainty Procedures (SWAT-CUP) software package (Abbaspour 2011). The SWAT-CUP offers a semiautomatic or combined manual/automatic calibration of SWAT projects, allowing the user to control the range of parameter perturbations in seeking to identify their optimum values. Parameters can range either by a percentage from their initial values or within predefined lower and upper bounds. The Sequential Uncertainty Fitting (SUF-2) algorithm (Abbaspour et al. 2007) was used in this study, which is the most efficient option for large regional applications (Schuel et al. 2008a, 2008b) and is highly recommended for the calibration of SWAT models (Arnold et al. 2012).
The SUFI-2 hydrologic calibration was followed by sediment and nutrient calibration, which was performed by manually adjusting the appropriate parameters. As previously mentioned, the management factor (USLE_P) of the MUSLE equation was the primary driving factor of controlling erosion simulation and sediment delivery to streams. River nutrient yields were calibrated based on several other parameters that govern nutrient soil availability and cycling. Some of them were the N and P percolation coefficients (NPERCO and PPERCO), the concentrations of organic forms of N and P in soil at the beginning of the simulation (SOL_ORGN and SOL_ORGP), as well as the coefficients governing denitrification (Neitsch et al. 2009). The goodness of fit criteria for all these variables, especially the percent bias and the correlation between observed and simulated loads were acceptable both at Grafton and other river locations across the basin and will be documented in a forthcoming study. It is beyond the scope of the current paper to present and discuss the detailed sediment and nutrient calibration approaches and results. Instead, two representative nutrient graphs of observed vs simulated loads at Grafton are included in figure 3.

**Table 1**

<table>
<thead>
<tr>
<th>Subbasin</th>
<th>Calibration point</th>
<th>US Geological Survey station</th>
<th>( r^2 )</th>
<th>Nash-Sutcliffe efficiency</th>
</tr>
</thead>
<tbody>
<tr>
<td>Minnesota</td>
<td>Jordan</td>
<td>5330000</td>
<td>0.79</td>
<td>0.69</td>
</tr>
<tr>
<td>St. Croix</td>
<td>St. Croix Falls</td>
<td>5340500</td>
<td>0.82</td>
<td>0.67</td>
</tr>
<tr>
<td>Chippewa</td>
<td>Durand</td>
<td>5369500</td>
<td>0.71</td>
<td>0.70</td>
</tr>
<tr>
<td>Wisconsin</td>
<td>Muscoda</td>
<td>5407000</td>
<td>0.64</td>
<td>0.61</td>
</tr>
<tr>
<td>Skunk</td>
<td>Augusta</td>
<td>5474000</td>
<td>0.91</td>
<td>0.89</td>
</tr>
<tr>
<td>Des Moines</td>
<td>Keosauqua</td>
<td>5490500</td>
<td>0.69</td>
<td>0.58</td>
</tr>
<tr>
<td>Illinois</td>
<td>Valley City</td>
<td>5586100</td>
<td>0.69</td>
<td>0.60</td>
</tr>
<tr>
<td>Iowa</td>
<td>Wapello</td>
<td>5465500</td>
<td>0.79</td>
<td>0.75</td>
</tr>
<tr>
<td>Royalton</td>
<td>Royalton</td>
<td>5267000</td>
<td>0.50</td>
<td>0.48</td>
</tr>
<tr>
<td>Rock</td>
<td>Joslin</td>
<td>5446500</td>
<td>0.70</td>
<td>0.59</td>
</tr>
<tr>
<td>Clinton</td>
<td>Clinton</td>
<td>5420500</td>
<td>0.65</td>
<td>0.53</td>
</tr>
<tr>
<td>Grafton</td>
<td>Grafton</td>
<td>5587450</td>
<td>0.73</td>
<td>0.70</td>
</tr>
</tbody>
</table>

**Figure 3**

Simulated versus observed flow time-series at (a) Clinton, Iowa, and (b) Grafton, Illinois; and (c) nitrate-nitrogen (NO\(_3\)), and (d) phosphorus (P) loads at Grafton along the Mississippi River for both the calibration and validation periods.
The atmospheric model solves its equations using spectral methods in the horizontal with a triangular truncation of 42 wavenumbers, which approximately corresponds to a horizontal grid of 2.8° latitude-longitude or around 230 km (143 mi) over the UMRB. The current climate for MIROC 3.2 is taken as the years 1981 to 1999 from its first ensemble member for the CMIP3 (LLNL 2013) “Climate of the 20th Century” simulations. These simulations include observed forcings of the major greenhouse gases (carbon dioxide [CO2], methane [CH4], and nitrous oxide [N2O]) are assumed to increase through the 21st century. An overview of the CMIP3 experiment design is given by Meehl et al. (2007).

For temperature and precipitation, we used monthly downscaled results of the medium-resolution MIROC 3.2 available from the “Downscaled CMIP3 and CMIP5 Climate and Hydrology Projections” archive (Downscaled CMIP3 and CMIP5 Climate and Hydrology Projections 2014). The downsampling method used was bias corrected with spatial disaggregation. In this procedure, biases were first removed from the model’s own climatology. Then, the departures from climatology for any given monthly time step were interpolated to a one-eighth degree latitude-longitude grid and superimposed on the observed climatology. The daily observed weather time-series were selected, formulated, and tested with SWAT by superimposing the resulting changes on the monthly weather generator statistics used in SWAT. The well-calibrated UMRB-SWAT model was then executed with new climatic information to perform the climate change impact assessment. We executed the model for 20 years for both the baseline (1981 to 2000) and the future climate (2046 to 2065).

**Agricultural Management Scenarios.**

Four agricultural management scenarios were selected, formulated, and tested with SWAT under the existing and future climate conditions in UMRB in order to compare their effects on pollutant losses from land to surface waters as well as their ability to sustain corn production. The implementation of these scenarios is of high interest in the Corn Belt, and these scenarios are investigated within the context of the Climate and Corn-based Cropping Systems Coordinated Agricultural Project (CSCAP 2013). The land use and cropping management scenarios included expansion of (1) C-C, (2) NT, (3) extended rotation of the typical C-S and C-C rotations with alfalfa (C-S-A-A-A), and (4) planting of rye as a winter cover crop in alternating years between row crop growing seasons in the C-S and C-C rotations. All scenarios were implemented entirely or to the greatest feasible extent for the UMRB. Table 2 summarizes implementation of these scenarios in SWAT.

**Results and Discussion**

**Water Balance under the Historical and Future Climate.** The calibrated SWAT-UMRB model was executed with the current (1981 to 2000) and future (2046 to 2065) climate data. On an average annual basis, precipitation was found to decrease to 829 mm (32.6 in) from the baseline value of 884 mm (34.8 in; 7.5% decrease) for the years 1981 to 2000, while snowfall was found to decrease to 21 mm (0.8 in) from the baseline value of 26 mm (1 in; 23% decrease). The mean annual temperature averaged across the whole UMRB was 8.2°C (47°F) for the current climate but increased to 12.2°C (54°F)
for the future climate. Together with the 1.1 MJ m$^{-2}$ yr$^{-1}$ solar radiation increase in the study area, one would expect a considerable increase in mean annual ET, but this was not produced in the SWAT results. We inferred that the increase in potential ET resulting from increased temperature and radiation roughly balanced the decrease of water availability on the ground due to reduced precipitation. SWAT simulated similar annual ET rates: a mean annual ET of 614 mm yr$^{-1}$ under the baseline climate and 603 mm yr$^{-1}$ for the future climate (approx. 24 in yr$^{-1}$). Spatial analysis across the basin (not presented here) showed that precipitation alterations in the future climate were maximum (up to 150 mm [6 in] reduction) in the most southern part of the basin (western Illinois and Missouri area), while reductions occurred in the more intensely cultivated areas of UMRB, which could increase water stress for crop production. Evapotranspiration did not consistently follow this pattern, but was increased during the first stages of crop growth (May and June) under the predicted climate, which in combination with reduced precipitation caused an overall runoff reduction until the end of summer. The alterations of the water balance components within the growth period, especially the first stages when fertilizers are applied, imply an assisting role of climate change in reducing pollutant losses from land to waters. On the other hand, reduced water availability may have a negative impact on crops, which may meet more water stress days.

**Pollutant Losses from Scenario Implementation.** In our calibrated SWAT-UMRB model, sediment losses from the agricultural land under the current climate were 0.95 t ha$^{-1}$ yr$^{-1}$ (1.0469 tn ac$^{-1}$ yr$^{-1}$) within the 20-year baseline period (1981 to 2000). Similarly, NO$_3$-N and total N (TN) losses in the current modeling study were predicted as 14.6 and 17 lb ac$^{-1}$ yr$^{-1}$, respectively. For TP, our

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**Table 2**

Management scenarios implemented in the Soil and Water Assessment Tool (SWAT) simulation of the agricultural land of the Upper Mississippi River Basin (UMRB).

<table>
<thead>
<tr>
<th>Scenario</th>
<th>Where implemented</th>
<th>How implemented in SWAT</th>
<th>Main purpose</th>
</tr>
</thead>
<tbody>
<tr>
<td>Continuous corn (C-C)</td>
<td>To all corn–soybean (C-S) rotations of the baseline in UMRB.</td>
<td>Changing soybean with corn and increasing N fertilization by 50 kg ha$^{-1}$ yr$^{-1}$.</td>
<td>Increase corn production in the long term.</td>
</tr>
<tr>
<td>No-tillage (NT)</td>
<td>To all C-S and continuous corn rotations with conventional, reduced, or mulch tillage.</td>
<td>Apply tillage passes with lower depth (25 mm) and low mixing efficiency (0.05) and reduce the crop factor (USLE_C) in the crop database. Reduce CN values and increase OV_N.</td>
<td>Reduce erosion, N, and P losses from fields to waters.</td>
</tr>
<tr>
<td>Extended rotation (C-S-A-A-A)</td>
<td>To all C-S and C-C rotations.</td>
<td>Apply the typical C-S and C-C rotations followed by 3 years of alfalfa cultivation. Reduce CN values (CNOP) in the third year when planting alfalfa, increasing them again when the first crop (corn or soybean) is planted. Fertilization of alfalfa with 20 kg P ha$^{-1}$ yr$^{-1}$ and 30% N fertilization reduction in corn.</td>
<td>Reduce erosion, N, and P losses from fields to waters.</td>
</tr>
<tr>
<td>Cover crop (rye)</td>
<td>To all C-S and C-C rotations in the UMRB.</td>
<td>Plant rye as a winter cover crop (October to April) between row crops in both the C-S and C-C rotations; the cover crop was grown in alternating years (e.g., between the first cycle of corn and soybean but not between the next cycle of soybean and corn, etc.)</td>
<td>Reduce erosion, N, and P losses from fields to waters.</td>
</tr>
</tbody>
</table>

**Notes:** N = nitrogen. P = phosphorus. CN = curve number for surface runoff simulation. CNOP = updated curve number due to a specific practice. OV_N = Manning’s roughness coefficient for overland flow (Neitsch et al. 2009).
mean annual prediction was 1.3 kg ha⁻¹ yr⁻¹ (1.1 lb ac⁻¹ yr⁻¹). It should be noted that upland erosion and nutrient outputs from agricultural fields were not directly measurable variables. Pollutant yields were measured and reported along streams and rivers, while the official USGS data corresponded to a lower total N and P load on a per hectare of the upstream area basis at Grafton, Illinois, compared to the upland pollution from agricultural fields analyzed by our results. This was mainly attributed to the unit area contribution of nonagricultural areas to water pollution, which was much lower than that of the agricultural land. The reliability of predictions from the agricultural land was based on the ability of SWAT to capture spatial heterogeneity given the accuracy of our model parameterization and the success of the calibration process. However, even though there is some uncertainty regarding the predicted absolute values, the purpose of the study was to analyze relative comparisons of the productivities and the susceptibility of the agricultural land in pollutant loss under various management and climatic conditions, which is a straightforward application of SWAT.

Figure 5 shows the mean annual sediment and P yields generated from the UMRB agricultural land for both the current and predicted climate and under the implementation of both the baseline management scenario and the four scenarios listed in table 2. These pollutants were closely linked in the UMRB-SWAT model as both organic, and the greatest mineral P pool were attached to soil particles and transported with them. The C-C scenario resulted in a slightly reduced sediment from HRUs compared with the baseline. Although corn was erodible to the same extent as soybean according to the attributes of both crops in SWAT (USLE_C factor, CN values), the replacement of soybean with corn produces higher residue amounts, resulting in reduced soil erosion. The expansion of NT was the most promising scenario, which resulted in drastic sediment and P load reduction from the agricultural land. Sediment reduction approached 70%, and P reduction was 40%. The smaller P reduction was attributed to mineral forms linked directly with runoff that were not influenced significantly by NT. The extended rotation (C-S-A-A-A) resulted in about 50% load reductions for both pollutants mainly due to decreased simulated erosion caused by alfalfa. For P, fertilization was slightly reduced compared to corn and soybean which contributed to some extent to the results produced. Finally, the establishment of rye as a winter cover crop within the traditional UMRB rotations (C-S or C-C) resulted in reduced sediment and TP loads of 25% to 30%, because of increased soil protection. Under the future climate, all scenarios behaved similarly, with both pollutants reduced by a further 20% to 25% compared with the baseline management conditions. Reduced precipitation and surface runoff were responsible for this result. All management scenarios behaved similarly when implemented under the climate change scenario. Pollution was reduced in the future climate, and various management scenarios further decreased pollutant loads.

Figure 6 presents the mean annual NO₃-N and TN yields estimated from the UMRB agricultural land with the historical and predicted climate and under the implementation of both the baseline management scenario and the four scenarios listed in table 2. The NO₃-N losses were highly governed by subsurface flow pathways (tile and baseflow) while the TN losses also included organic N attached to sediment. The C-C scenario resulted in increased NO₃-N and TN loads compared to the baseline. Increased N fertilization of 50 kg N ha⁻¹ y⁻¹ (44 lb ac⁻¹ yr⁻¹) during each year of C-C cultivation, relative to the N application rates to corn in the C-S baseline rotation, was responsible for this increase. The NO₃-N losses for the C-C scenario increased by 6.7 kg ha⁻¹ y⁻¹ (5.9 lb ac⁻¹ yr⁻¹) or 45% from the baseline due to higher N availability in soil and subsoil. Nitrogen transported with sediment, however, was reduced to some extent due to increased residue levels that resulted in less erosion and sediment transport. This can be inferred from the increase in TN yields (figure 6), which was 6 kg ha⁻¹ y⁻¹ (5.3 lb ac⁻¹ yr⁻¹), which was less than the corresponding NO₃-N increase of 6.7 kg ha⁻¹ y⁻¹ (6 lb ac⁻¹ yr⁻¹). Adoption of NT had little effect on NO₃-N losses. The model simulated a very slight increase in NO₃-N rate showing slightly increased N availability in soil and subsoil.
was considerably reduced. The C-S-A-A-A-A rotation was the most effective management scenario because it resulted in 50% or more reductions of both NO₃-N and TN on a mean annual basis. The reason was not only the reduced erosion but also the absence of N fertilization in alfalfa. Cover crops had a significant effect on N pollution mitigation as well. Reductions were on the level of 25% for both NO₃-N and TN due to N uptake and erosion protection during periods when soils were susceptible to pollutant transport. All four scenarios produced similar results when applied under the future climate. Pollutant levels were consistently lower compared to the respective loads under the historical climate due to the water balance alterations in the future climate. Similarly with P, the future climate scenario resulted in reduced N from the entire agricultural land of UMRB on a mean annual basis.

**Predicted Yields from Scenario Implementation.** Figure 7 summarizes the SWAT crop yield results for corn and soybean, while their interannual variability is presented in table 3. Mean annual simulated corn and soybean yields in the baseline scenario were 8.9 and 2.9 t ha⁻¹ y⁻¹ (3.56 and 1.16 tn ac⁻¹ yr⁻¹) respectively across the agricultural land of the UMRB. A slightly increased average annual corn yield of 9.3 t ha⁻¹ y⁻¹ (3.72 tn ac⁻¹ yr⁻¹) occurred for the C-C scenario, due to the increased N fertilization. However, under this management scenario corn average yield was calculated from a 20-year yield population, which may have had some statistical impact, because the corn production years were double those simulated in the baseline (10 years of corn). On the other hand, NT applied in all C-S and C-C HRUs of UMRB did not have any impacts on yield. The results can however be considered promising as the practice was able to sustain yields under the new residue management conditions. Finally, reduced crop yields were predicted for both the C-S-A-A-A and cover crop scenarios, which were close to 5% for both corn and soybean. These results showed that SWAT was sensitive to N fertilization even though corn was preceded by three years of alfalfa, which resulted in increased N availability in soil. In addition, it is possible that the corn and soybean yields were affected by the climatic conditions in the specific years that both crops were grown in the C-S-A-A-A-A rotation. This may have caused undesirable effects on the annual simulated growth if these years were among the driest in large parts of the basin. Alfalfa yield was simulated as 7.5 t ha⁻¹ y⁻¹ (8.265 tn ac⁻¹ yr⁻¹) of alfalfa production and could have again been affected by the climatic conditions during the years that the simulated alfalfa growth occurred. Substantial market changes would have to occur in order for expanded alfalfa production to emerge as a viable crop alternative in the UMRB region.

Finally, corn and soybean yields were also reduced when rye was grown as a winter cover crop. Climate variability effects between years was not present here as the typical C-S, S-C, and C-C rotations of the baseline were applied throughout the 20-year period. Therefore, the slight reductions (2% to 6%) were attributed to reduced nutrient availability for the subsequent crop due to uptake by rye. Simulated rye yields ranged from 0.2 t ha⁻¹ y⁻¹ (0.2204 tn ac⁻¹ yr⁻¹) in the driest in large parts of the basin. Alfalfa growth occurred. Substantial market changes would have to occur in order for expanded alfalfa production to emerge as a viable crop alternative in the UMRB region. Expanded rye yields ranged from 0.2 t ha⁻¹ y⁻¹ (0.2204 tn ac⁻¹ yr⁻¹) in the driest in large parts of the basin.

The predicted corn and soybean yields consistently declined for all four scenarios with reference to the same management conditions of the current climate. The ranking of practices in terms of their influence on yields was the same, with the extended rotation having the highest negative impacts. The reduced precipitation predicted for the future climate period was the main reason for the lower crop yields, due to increased water stress.

The conclusions drawn here were based on an analysis of water quantity and quality variables at the large basin scale. It would be useful to analyze the results by mapping the effectiveness of each scenario in reducing pollution and in sustaining crop yields at the 12-digit subbasin level. However, improved representation of existing conservation practices, nutrient application rates, and other management practice aspects are needed in order to simulate accurate combinations of practices across specific landscapes. Moreover, a next step would be to translate upland pollutant reduction from agricul-
ture to precise river pollution levels, which can more closely address the level of pollution in the Gulf of Mexico hypoxic zone. A practice allocation across the landscape of UMRB would also require a clear cost estimation of the practices in different locations. In addition, incorporation of HRUs within the 12-digit subwatersheds is needed to better represent the impacts of different combinations of cropland landscapes and management practices.

This study was based on a future climate projection from a single GCM, the MIROC 3.2 model. A next step in better addressing climate change in the UMRB would be to examine SWAT results using climate projections from an ensemble of different GCMs rather than a single GCM. This could better define the uncertainty of future climate predictions and their impact on SWAT predictions.

Summary and Conclusions
This study examined the impact of four agricultural management scenarios in the intensely row-cropped UMRB region for both current and future climate conditions and a climate change projection generated with the MIROC 3.2 model for a future midcentury time period (2046 to 2065). The major effects of the climate projection resulted from reduced precipitation and increased temperature occurring within the crop-growth cycles, especially the first stages of crop development. This reduced soil water availability and runoff potential had a positive effect on pollution mitigation but a slightly negative impact on crop production. All four scenarios behaved similarly under the current and future climate resulting in reduced erosion and nutrient loadings to surface water bodies. Increased N pollution was predicted only for the C-C scenario. No-till was the most environmentally effective scenario and was predicted to sustain crop production levels similar to baseline crop yield estimates. The effective extended five-year rotation of the typical crops with alfalfa seemed to be the least economically desirable, given the assumption that alfalfa was less valuable than the other crops. The establishment of rye as a winter cover crop was also effective in reducing erosion and both sediment-bound and soluble forms of nutrients with a small sacrifice in crop yields and would be expected to be even more effective if grown within every winter period, rather than alternating growing periods as simulated in this study. The trend of the simulated effects of the scenarios tested was in agreement with findings from several experiments that were recently reported by the Iowa Department of Agriculture and Land Stewardship (IDALS 2013).

This study highlights the capabilities of SWAT in connecting agricultural management strategies with hydrologic-process simulations at the river basin scale. It also supports its use as a component of an integrated decision support system for the complex Corn Belt agricultural systems. Such tools can provide scientifically based estimates of the effect of a wide array of alternative cropping and management strategies under different climatic conditions, enabling informed choices and affecting environmental and economic sustainability of the region in the coming decades.

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References
Figure 7
Average annual corn and soybean yields for the years when the crops are growing in Upper Mississippi River Basin with the historical and future climate under the baseline and four agricultural management scenarios: continuous corn (CC), no-till, extended rotation (C-S-A-A-A), and cover crop. Corn and soybean in the C-S-A-A-A scenarios grow once every five years, thus their average annual production within the entire simulation period is calculated by diving yields by five.

Table 3
Analysis of Soil and Water Assessment Tool simulated yields in the Upper Mississippi River Basin for the period 1981 to 2000 or 2046 to 2065 (General Circulation Model scenario). CC = continuous corn. GCM = General Circulation Model.


IDALS (Iowa Department of Agriculture and Land Stewardship). 2013. Iowa Nutrient Reduction Strategy: A science and technology-based framework to assess and reduce nutrients to Iowa waters and the Gulf of Mexico. Iowa Department of Agriculture and Land Stewardship (IDALS), Des Moines, IA; Iowa Department of Natural Resources (IDNR), Des Moines, IA; and Iowa State University College of Agriculture and Life Sciences (ISUCALS), Ames, IA. http://www.nutrientstrategy.iastate.edu/documents.


